

Biodegradation of Environmental Chemicals Pollutants: Evolutionary Insights

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Abstract – An increase in organic pollutant is a major concern in the contemporary world. Organic pollutants may be treated through physical and chemical processes, but these processes are toxic and not environmental friendly because their final product may still remain toxic till the very end. Hence, the biological approaches may be a suitable alternative towards bioremediation practices being not only cost effective but eco-friendly as well. Moreover, the final product happens to be less toxic as compared to other approaches. The microorganisms and plants (bioremediation) are used to remediate the polluted environments widely and is emerging as a promising and appealing area of environmental biotechnology. Apart from using the whole cell microorganisms, the use of their extracellular and/or cell-free enzymes has been advocated as an innovative technique to abate pollution. Employment of extracellular enzyme for the removal of pollutants has several advantages over using whole microbial cell. The present review attempts a brief survey of many aspects dealing with the characteristics and potential abilities of both cell-present and cell free extra cellular enzymes in bioremediation of various organic pollutants.

Keywords: Bioremediation, Organic Pollutant, Microbial Enzymes, Extracellular Enzymes, etc.

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INTRODUCTION

Albeit a few yeasts, green growth, diatoms and in addition higher plants and creatures use an assortment of chemicals, their exercises are regularly unobtrusive in comparison to the changes affected by heterotrophic bacterial and parasites dwelling in a similar living space. Along these lines, the bacterial and contagious populaces show in soil and water are the main operators for biodegradation of the natural chemicals.

Biphenyl and chlorinated biphenyls are widely utilized as a part of industry and horticulture and are known to be dangerous ecological contaminations. In comparison with the abundance of data accessible on the biodegrade at particle of different classes of fragrant mixes significantly less I's thought about the biodegradation of biphenyl and its subsidiaries m diverse living beings. The pathway for the corruption of biphenyl mixes has been examined primarily in the microscopic organisms of sort *Pseudomonas*. Be that as it may, the capacity of *Micrococcus* species to corrupt biphenyl has not been accounted for. Nevertheless more concentrated in other bacterial frameworks are important with a specific end goal to assess the all-inclusiveness and transformative essentialness of the biodegradative pathways. Therefore, the present examination was attempted to explain the biodegradative pathway of biphenyl and

4-chlorobiphenyl by a *Miorococons* sp, confined from soil by advancement culture.

The late 1800s and early 1900, has witnessed a dramatic increase in the range of chemically synthesized products which include pesticides, plastics, hydrocarbon fuels, soaps, detergents and other useful substance. The effects of these chemical substances on the environment are a consequence of a sequence of processes that depend on the properties of individual chemical. Halogenated organic pollutants (HOPs), such as polychlorinated biphenyls (PCBs), dichlorodiphenyltrichloroethane (DDT), polybrominated diphenyl ethers (PBDEs), dechlorane plus (DP), and decabromodiphenyl ethane (DBDPE), have been of great concern due to their persistence, bioaccumulation and potential toxicity to wildlife and human. PCBs were used primarily as dielectric and coolant fluids in capacitors, transformers and. electric fluids PBDEs (including Penta-, Octaand Deca-BDE commercial formulations), DP and DBDPE are some widely used flame retardants in electronics, textiles, and thermoplastics, polyurethane.

BIODEGRADATION OF BENZOIC CORROSIVE AND ITS SUBSIDIARIES:

Benzoic corrosive: The bacterial corruption of benzoic corrosive happened regularly by an underlying twofold hydroxylation took after by ring cleavage, in spite of the fact that in a couple of examples monohydroxy subsidiaries were likewise hypothesized as intermediates *cis*-3,5-aCyclohexadiene-1,2-diol-1-carboxylic corrosive was appeared to be a middle of the road in the transformation of benzoate to catechol by *Acetobacter caliginis* and *Pseudomonas* SP (Reiner, 1971; Knackmuss and Reineke, 1973) Yamaguchi and Fujisawa (1982) filtered benzoate-1, 2-dioxygenase from *Pseudomonas aeruginosa*. The compound was made out of a NADH-Cytochrome *c*-reductase and a benzoate oxygenase, which required Fe^{2+} and NADH for maximum activity. The dihydrodiol dehydrogenase catalyzing the change of *cis*-diol to catechol has been cloned from *A. eutrophus*.

In few instances, benzoate was degraded by the initial formation of salicylate or 3-hydroxybenzoate or 4-hydroxybenzoate. Most of the organisms degraded benzoate through monohydroxylation response to yield 4-hydroxybenzoate. The monohydroxy benzoic acids are further metabolized through the presentation of another hydroxy group at the ortho or para position to the current hydroxy group.

Salicylic corrosive: Salicylic corrosive was oxidatively decarboxylated to catechol by a few types of microorganisms having a place with family *Pseudomonas*, *Mycobacterium fortuitum* (Tsukamura, 1965), *Acetobacter eullulariapullulans* and a few types of bacilli. Salicylate-1-monooxygenase (decarboxylating), an outer flavoprotein monooxygenase has been cloned from *Pseudomonas putida*. This chemical has been examined widely and proposed an instrument for hydroxylation response. Growths in general appear to hydroxylate salicylate at the third position yielding 2,3-dihydroxybenzoic corrosive. In spite of the fact that this pathway was demonstrated in *A. niger*, *A. nidulans* and *Trichoderma lignorum*, the protein catalyzing this response was not acquired in *in vitro* cell frameworks. Salicylate was changed over to gentisic corrosive in *Pseudomonas*, *Lignobacter* and a few types of bacilli.

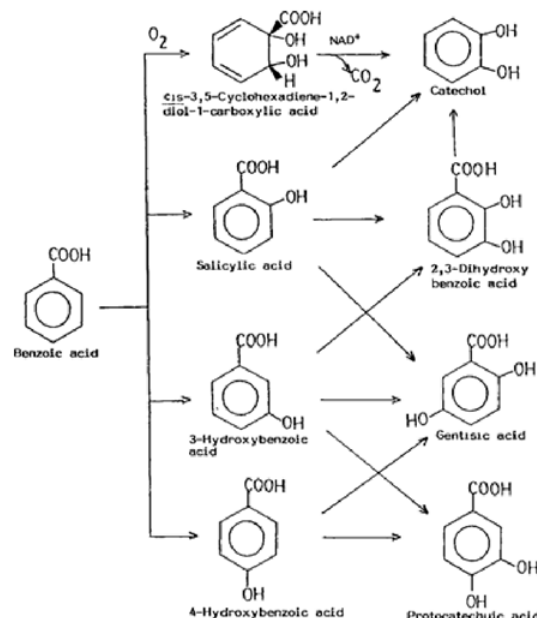


Figure-1

The degradation of 3-hydroxybenzoic corrosive continues through hydroxylation at the 4th or 6th position yielding protocatechuic or gentisic individually. 3-Hydroxybenzoate was hydroxylated to protocatechuic by *P. testosteroni*, *A. japonicus* and *A. niger*. A few *Pseudomonas* species and *Streptomyces* changed over 3-hydroxybenzoate to gentisic corrosive. Both these pathways were appeared in a few types of bacilli. A mutant strain of *testosterone* collected 2,3-dihydroxybenzoic corrosive from 3-hydroxybenzoate as a dead-end metabolite. As of late, Starovoirov *et al.* (1985) revealed the catabolism of 3-hydroxybenzoate by *P. putida* BS893 by another pathway by means of 2,3-dihydroxybenzoic corrosive and catechol. This life form was likewise appeared to change over 3-hydroxybenzoate to gentisic and protocatechuic. The different pathways for the corruption of benzoic corrosive and monohydroxybenzoic acids are appeared in Fig 1.

BIODEGRADATION OF HALOGENATED SWEET-SMELLING MIXES:

Halogenated sweet-smelling mixes (chloroaromatics in specific) are utilized as a part of vast amounts in an assortment of mechanical and horticultural applications, for example, bug sprays, herbicides, plasticizers, solvents and others. These are the most persevering environment toxins due to their impervious to microbial corruption and danger. In any case, ongoing investigations have uncovered that haloaromatic mixes are biodegradable.

Halogenated benzoic acids: Microorganisms have been accounted for to corrupt halobenzoates by one of two unmistakable courses. These are dehalogenation before ring cleavage and dehalogenation in the wake of ring cleavage. A few

bacterial and parasitic strains have been appeared to catabolize chlorobenzoates by dehalogenation before ring cleavage. While 3-chlorobenzoate was converted to 3-hydroxybenzoate by a soil *Pseudomonas*, 4-chlorobenzoate was converted to 4-hydroxybenzoate in *Arthrobacter*, *Koccardia* sp., and *Aspergillus* niger. The cometabolism of 3-chlorobenzoate and 3-fluorobenzoate was exhibited in *Vibrio* sp., *Nocardia* sp. and in a few blended societies. The greater part of creatures equipped for development on halobenzoates catabolize these mixes by means of the comparing halocatechols, and dehalogenation happens in the wake of ring cleavage.

Subsequently, the oxidation of fluorobenzoates and chlorobenzoates to fluorocatechols and chlorocatechols separately, have been appeared to happen in *Pseudomonas* sp.), *A. eutr.Q-phus* and *Acinetobacter* tercalcoacetilCUS. A considerable lot of the compounds catalyzing these responses prompting dehalogenation more likely than not casual substrate specificities or evolutionarily specialized catalytic functions. Microorganisms performing such dehalogenation must be safe to the first halogenated compound as well as to the halogenated intermediates. Many of the qualities engaged with the degradation of chloroaromatics are plasmid-borne. The nearness of the plasmid pAC25 in *Pseudomonas putida* is in charge of the usage of 3-chlorobenzoate. The conversion of 3-chlorobenzoate to 3-chlorocatechol most likely catalyzed by the compounds of benzoate corruption. Plasmid encoded chemicals are in charge of the dissimilation of the subsequent chlorocatechols. Pyrocatechase II particularly follows up on chlorocatechol and this chemical did not follow up on catechol. As opposed to benzoate corruption which continued by the ortho-cleavage of catechol, the plasmid indicated pathway for chlorocatechol continued through maleylacetate as a transitional. The different pathways for the metabolism of halogenated benzoic acids are demonstrated in Fig 2.

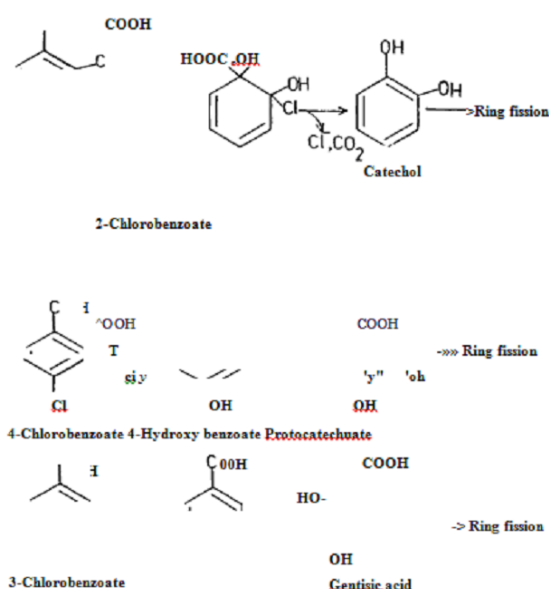
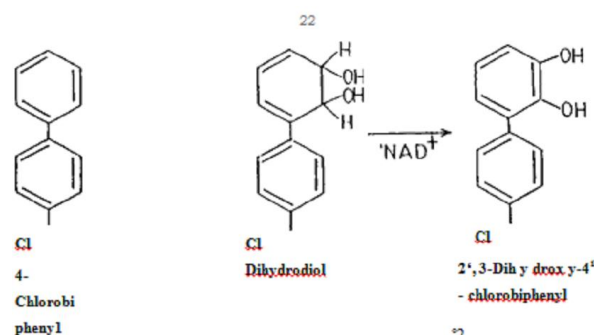


Fig. 2. Pathway for the biodegradation of halogenated benzoic acids

Halogenated biphenyls: Polychlorinated biphenyls (PCBs) are overall contaminations due to their hydrophobicity and protection from organic corruption. Ongoing investigations have demonstrated that few unadulterated and blended culture of microorganisms are equipped for utilizing the mono-, di-, and trichlorinated biphenyls. A few examiners have considered the debasement of PCBs by the microbes of the genera *Pseudomonas*, *A. chromobacter*, *A. caligenes*, *Acinetobacter* and *Rhizopus*. Notwithstanding, there are just couple of reports of microscopic organisms fit for corrupting the all the more very chlorinated PCBs.



The significant pathway for the bacterial debasement of chlorinated biphenyls includes the development of 2,3-dihydroxy subsidiary which experiences 1,2-dioxygenative ring cleavage to produce the comparing chlorobenzoates. A large portion of the bacterial strains corrupted 4-chlorobiphenyl to 4-chlorobenzoate through 2,3-dihydroxy-4-chlorobiphenyl by a meta-cleavage pathway as appeared in Fig 3. There are additionally reports of corruption of PCBs by new course in *Alcaligenes*, *Eutropha* and in a *Pseudomonas* sp. which delivered

chloroacetophenones and different metabolites. Be that as it may, the enzymological investigations of this pathway has not been done numerous microorganisms have been appeared to cometabolize polychlorinated biphenyls to the related chlorobenzoates.

FRAGRANT RING CLEAVAGE PATHWAYS:

The oxidative cleavage of the fragrant ring catalyzed by the dioxygenases is the most basic advance in the mineralization of sweet-smelling mixes as the aliphatic mixes shaped by the activity of dioxygenases enter the TCA cycle by straightforward decarboxylation, hydrolysis and isomerization responses. Dioxygenases cut the sweet-smelling ring containing two hydroxyl bunches that are either ortho or para to each other. There are two particular methods of oxidative cleavage of the fragrant ring to be specific *QrlllQ*- and meta cleavages. Cleavage of the bond between two contiguous carbon atoms that convey hydroxyl bunches is known as "ortho" or then again "intradiol" cleavage and the pathway by which the results of such cleavage are metabolized is known as the ortho or *O*-ketoadipate pathway. Cleavage of the bond between two carbon atoms, just one of which carries a hydroxyl gathering, the other carbon atom being either unsubstituted or substituted with other than a hydroxyl group. This type of cleavage is known as "meta" or "extradiol" cleavage and the pathway by which items of such cleavage are metabolized is called the meta-pathway. When the hydroxyl groups are para to each other as in gentisic corrosive (2,5- dihydroxybenzoic corrosive), oxidative cleavage is catalyzed by gentisate-1,2-dioxygenase, and the ensuing pathway is the gentisate pathway. Host of the ring cleavage dioxygenases have been appeared to contain nonheme press as the sole cofactor with the exception of the homoprotocatechuate-2,3-dioxygenase from *Bacillus brevis* which was appeared to contain manganese II. While the majority of the meta-cleavage dioxygenases have been appeared to contain ferrous iron (Fe^{2+}), ortho - cleavage dioxygenases contain ferric iron (Fe^{3+}).

Organic pollutants-Types and properties: Organic pollutants are chemical compounds that contain carbon and have a demonstrably negative effect on one or more components of the environment. Organic pollutant can be placed into three general classes: (i) hydrocarbons, (ii) oxygen, nitrogen and phosphorus compounds and (iii) organometallic compounds. The major category of organic pollutants includes the hydrocarbons and related compounds, which contains such compounds as Dichloro Diphenyl Trichloroethane (DDT), the dioxins and the polycyclic aromatic hydrocarbons (PAHs). These compounds contain the elements of carbon and hydrogen, with some containing chlorine and oxygen as well. There are a limited number of types of chemical bonds present, which are principally C-H, C-C, C-Cl, C=C and C=C (aromatic). All of these

bonds are relatively stable and have limited polarity and this property is then conferred onto the related compounds. (1) Owing to low polarity, hydrocarbons, in general are lipophilic, poorly soluble in water and persistent in the environment. This class includes the most toxic organic compound, abiotic in origin: 2, 3, 7, 8-tetrachlorodibenzo (1, 4) dioxin, also known as 2, 3, 7, 8-TCDD or [7] TCDD.

DEGRADATION OF ORGANIC POLLUTANTS:

Organic chemicals that are introduced into the environment are subjected to various physical, chemical, and biological processes which act in an interconnected way in environmental systems to determine the overall fate of the compound. The neutralization when done through chemical means, a huge amount of acid is used, which is neither economical, nor safe and poses serious health [14] hazards. There are many processes for the degradation of organic pollutants. Some processes for degradation of organic pollutant are listed below:

1. **Physical processes:** Physical processes have been used for the degradation of organic pollutant from many decades, which may include various processes like photocatalytic degradation by using Ag-modified Zn GeO nanorods, TiO₂/graphene oxide 2 4 2 nanocomposite hydrogels, Bio-silica coated with amorphous manganese oxide etc. Decomposition of these organic pollutants via catalytic/ photocatalytic oxidation is considered to be the most efficient green method for organic waste management. Visible-light response semiconductors have attracted interest of many researchers as the efficient photo catalysts. There are many catalysts used for photocatalytic degradation of organic pollutants. TiO₂ used as a 2 photo catalyst because of its low cost, chemical stability, non-toxicity. TiO₂ is preferred because 2 it is a promising photo-oxidation catalyst and has strong oxidizing ability of photo-induced holes. Many researchers coupled TiO₂ with narrow band 2 gap semiconductors which enhanced the separation of photo induced charges by formation of heterogenous junctions. Researchers modified the surface of Ag PO using TiO₂ by sol gel process. Researches in their study deposited Ag₃PO₄ nanoparticles onto TiO₂ to form heterostructure.
2. Researchers prepared Ag/Ag₃ PO₄ /TiO₂ heterostructure photo electrodes using a sequential chemical deposition and UV-reduction method. One of the effective methods was using TiO₂ nanoparticles as a photo catalyst for the degradation of organic compounds due to their non-toxicity, low

cost, physical and chemical stability, and high reactivity.

Chemical Processes: The chemical methods for bioremediation include electrochemical dehalogenation of chlorinated benzenes, in this the chlorine is eliminated step by step from the highly chlorinated benzenes to yield less-chlorinated benzenes and finally transform to benzene. It was analyzed that chlorobenzenes and the cathodic reaction pathway for hexachlorobenzenes as follows: hexachlorobenzene pentachlorobenzene 1, 2, 3, 5-tetrachlorobenzene 1, 2, 4-trichlorobenzene 1, 4-dichlorobenzene monochlorobenzene benzene. The catalytic degradation of organic molecules through MnO nanostructures are 2 concerned, several groups reported the mineralization of various organic compounds/ dyes, such as Rhodamine B (RhB), MB, Benzyl alcohol (BA) in the presence of strong oxidizing agents at elevated temperature.

Biological processes: Bioremediation of organic pollutant contaminated soil offers a cost-competitive treatment for many sites that are currently facing costly incineration or the extended liability of land disposal. In the field, under conditions of full-scale site remediation, this [32, technology has been shown to be cost effective 33]. Different types of biological processes include bio-attenuation, bio-stimulation and bio augmentation.

CONCLUSION:

Organic pollutant contaminated sites treatment through bioremediation is best because the approach is not only sustainable but eco-friendly. Moreover, there is no collateral loss of habitat quality. The immense biodiversity of microorganism in our environment offers greater potential for transformation of toxic compounds to less toxic by-products. Microbial enzymes play a major and crucial role in biodegradation of soil contaminated with organic pollutants such as diesel, petrol or PAHs compound etc. In future, these enzymes open further exploration by researchers which is likely to open a new era of microbiology aiding various environments clean up technologies. In this way our examination has helped to know the debasement pathway of biphenyl and 4-chlorobiphenyl in *Micrococcus* sp. There was an entire mineralization of biphenyl by the disconnected bacterium. 4-chlorobiphenyl was debased to 4-chlorobenzoic corrosive, which can be used further by soil microbes. A dioxygenase catalyzing the meta-ring cleavage of 2,3-dihydroxybiphenyl was mostly refined and examined a portion of its properties.

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